

## EXPLORING LAND COVER EFFECTS ON URBAN AIR QUALITY: A CASE OF 659 DISTRICTS IN INDIA

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### ABSTRACT:

Land use and land cover changes (LUCC) affects the atmospheric environment directly or indirectly. Therefore, understanding the atmospheric response to LUCC is of great significance to maintain and improve the ecological environment. In this study, based on fine particulate matter (PM<sub>2.5</sub>) and LC products, we first compared the differences of PM<sub>2.5</sub> between urban and surrounding areas, and then further investigated the variations of PM<sub>2.5</sub> in different and land cover (LC) using Mann-Kendall (MK) test and Sen's trend analysis approach at the district-level in India during 1998-2015. The results showed that the numbers of districts where the differences of PM<sub>2.5</sub> ( $D_{PM_{2.5}}$ ) between urban and the surrounding areas were greater than zero were increasing during 1998-2015. There is an upward tendency of annual mean PM<sub>2.5</sub>. The annual mean PM<sub>2.5</sub> was higher than 40 $\mu\text{g}/\text{m}^3$  in 58% of India's areas where there were mainly located in the Ganges plains of northern India with cropland (L01) and urban areas (L07). The annual mean PM<sub>2.5</sub> was less than 10 $\mu\text{g}/\text{m}^3$  were mainly found in north-western India with permanent ice and snow (L10), accounting for 10% of India's area. There are significant positive trends of PM<sub>2.5</sub> concentration in 90% of cropland (L01) and 88% of urban area (L07) and the average slope were 0.83 $\mu\text{g}/\text{m}^3$  and 0.82 $\mu\text{g}/\text{m}^3$  respectively, which were higher than those in the rest of LC. This research serves as the basis of reference for the equitable allocation of land resources and restructuring of land use and land cover patterns in urban areas of India that severely affected by air pollution.

### 1. INTRODUCTION

The remarkable conversion of land from natural and agricultural areas into residential and urban have taken place since the rapid development of urbanization. These severe unsustainable use of natural resources over the years may lead to tremendous environment degradation (Justice et al., 2015). Deforestation, grassland degradation, soil desertification, together with other urbanization process impacts, all may reduce the air purification capacity of vegetation around the city. Furthermore, it have a significant positive effect on national fine particulate matter (PM<sub>2.5</sub>) concentrations increase in developing countries (Wang et al., 2019). Therefore, a better understanding of land use and land cover changes (LUCC) and their interactions with the atmospheric environment is essential for the sustainable management of natural resources, environmental protection and air quality, especially for address these LUCC issues that associated with air pollution (Vadrevu et al., 2017).

As the world's most populous nation, more than 34% of the population of India live in urban areas, and it is projected to reach 52.8% by 2050 (United Nations, 2019). India has been progressing on a path characterized by rapid urbanization along with population growth. Over recent decades, India has witnessed increasing concentrations of PM<sub>2.5</sub> from multiple anthropogenic emissions sources, e.g., vehicles, manufacturing, electricity generation, construction and road dust, waste burning,

and household energy use (Kumar et al., 2017; Sharma et al., 2016). Especially, according to global urban ambient air quality database (WHO, 2016), there are 14 of world's 20 most polluted cities in India in terms of PM<sub>2.5</sub> levels. Rapid urbanization and population growth especially in the last decade have adversely affected urban climate and air quality of India (Saikawa et al., 2017; Sahu et al., 2017). Shi et al. (2017) founded that India consistently showed the largest PM<sub>2.5</sub> concentrations during 1999–2014, which were significantly higher than for any other countries. Han et al. (2015) analyzed PM<sub>2.5</sub> concentration varies for different land covers in China based on a single-year LC data. Li et al. (2018) explored the urban PM<sub>2.5</sub> pollution situation for 2014–2016 and investigated the impact of landscape factors on urban PM<sub>2.5</sub> in China at the city level. To date, many studies about the impact of changes in land use and land cover on air quality have been undertaken in China. However, only a few studies have investigated PM<sub>2.5</sub> concentration variation different land covers with long-term trends and spatial variations in India, where LUCC are also rapidly evolving in the process of urbanization. It is noted that these studies have limited applicability with regard to India, as a result of the large differences between India and the other study areas.

Therefore, based on longer-term LC products, we aimed at quantifying differences compared (PM<sub>2.5</sub>) concentrations in 659 districts of Indian urban and the surrounding regions during 1998–2015 in this study. Additionally, we explored how PM<sub>2.5</sub>

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concentration varies for different land covers using Mann-Kendall (MK) test and Sen's trend analysis approach.

## 2. MATERIALS AND METHODS

### 2.1 Land Cover

European space agency (ESA) released a consistent multi-temporal global land cover maps at 300 m spatial resolution covering 1992 to 2015 (<https://www.esa-landcover-cci.org/?q=node/164>), and the detailed calculation process can refer to the Plummer et al. (2017). In this study, it was employed to analyze the difference of PM<sub>2.5</sub> in the urban area (urban area LC class) and surrounding regions, and the original LC classes were further reclassified into 10 major LC types: agriculture (L01), forest (L02), mosaic herbaceous cover (L03), shrubland (L04), grassland (L05), sparse vegetation (L06), urban area (L07), bare area (L08), water (L09), and permanent snow and ice (L10).

### 2.2 PM<sub>2.5</sub> Concentration Differences

For evaluating the long-term PM<sub>2.5</sub> variabilities across India, this study used a satellite-based gridded PM<sub>2.5</sub> dataset obtained from the Dalhousie University Atmospheric Composition Analysis Group Work (available at [http://fizz.phys.dal.ca/~atmos/martin/?page\\_id=140](http://fizz.phys.dal.ca/~atmos/martin/?page_id=140)). This dataset is compiled using an integrated geophysical-statistical method, which provides a 0.01° × 0.01° grid at global scale for each year from 1998 to 2016 (van Donkelaar et al., 2016). The accuracy of this dataset is assessed by world-wide ground-measured PM<sub>2.5</sub> records in 2010, and the result presents a consistent agreement with ground-measured PM<sub>2.5</sub>, with an out-of-sample cross-validated R<sup>2</sup> value of 0.81 (N = 1855). The results indicate that the dataset can be utilized for large-scale PM<sub>2.5</sub> related studies. In this research, we applied the dataset from 1998 to 2015 to analyze the long-term spatial variation of PM<sub>2.5</sub> concentration in different LC of India. Additionally, the annual air quality standard of World Health Organization's (WHO) was used in this research (WHO, 2006). The standard has four levels: the air quality guideline (AQG; 10 µg/m<sup>3</sup>), and three interim targets (IT-1: 35 µg/m<sup>3</sup>; IT-2: 25 µg/m<sup>3</sup>; IT-3: 15 µg/m<sup>3</sup>). Especially, 40 µg/m<sup>3</sup> is also set as a reference value to evaluate the serious degree of local air pollution according to National Ambient Air Quality Annual Standards of India (NAAQAS).

To explore the temporal and spatial distribution characteristics of PM<sub>2.5</sub> concentrations between urban and the surrounding regions, the PM<sub>2.5</sub> concentration differences approach was employed in this study (Han et al., 2014). Firstly, average PM<sub>2.5</sub> concentrations were calculated at the municipal level using the PM<sub>2.5</sub> concentration dataset. The PM<sub>2.5</sub> concentration in the urban/nonurban areas (UrbanPM<sub>2.5</sub>/NonUrbanPM<sub>2.5</sub>) was then calculated based on the ESA LC dataset. The differences in PM<sub>2.5</sub> concentration ( $D_{PM_{2.5}}$ ) between UrbanPM<sub>2.5</sub> and NonUrbanPM<sub>2.5</sub> were obtained with the following equation:

$$D_{PM_{2.5}} = \text{UrbanPM}_{2.5} - \text{NonUrbanPM}_{2.5} \quad (1)$$

### 2.3 Trend Analysis Approach

Mann-Kendall (MK) trend test is a nonparametric test method (Mann., 1945; Kendall., 1948). This method does not require a normal distribution of data, and is an efficient

statistical tool for analyzing changes within long-term trends of air pollutant concentration data (Faridi et al., 2018; Bigi et al., 2014). The null hypothesis H<sub>0</sub> is that the data series  $x_k$  ( $k=1, 2, 3, \dots, n$ ) are independent from one another and has the same distribution, and the alternative hypothesis H<sub>1</sub> is that there is a monotonic trend in the data series. The MK trend test is calculated as following:

$$S = \sum_{i=1}^{n-1} \sum_{j=i+1}^n \text{sgn}(x_j - x_i) \quad (2)$$

where  $x_j$  is the sequential data value  
 $n$  is the size of the dataset  
 $\text{sgn}$  is calculated as follows:

$$\text{sgn}(x_j - x_i) = \begin{cases} 1 & \text{if } x_j > x_i \\ 0 & \text{if } x_j = x_i \\ -1 & \text{if } x_j < x_i \end{cases} \quad (3)$$

According to Mann (1945) and Kendall (1948), when  $n \geq 8$ , the test statistics  $S$  is approximately normally distributed with the mean and variance as follows:

$$E(S) = 0 \quad (4)$$

$$V(S) = \frac{n(n-1)(2n+5) - \sum_{m=1}^n t_m m(m-1)(2m+5)}{18} \quad (5)$$

where  $t_m$  is the number of extent  $m$

The standardized test statistics  $Z$  is calculated using the following formula:

$$Z = \begin{cases} \frac{S-1}{\sqrt{V(S)}} & S > 0 \\ 0 & S = 0 \\ \frac{S+1}{\sqrt{V(S)}} & S < 0 \end{cases} \quad (6)$$

$|Z\alpha| = 1.65, 1.96, \text{ and } 2.58$ , which correspond to the critical values at the significance level  $P=0.1, 0.05, \text{ and } 0.01$ , respectively. If  $|Z| > |Z\alpha|$ , the null hypothesis H<sub>0</sub> is rejected. In our study, annual averages of PM<sub>2.5</sub> concentrations were used for the long-term trend analysis. However, MK test indicates only the direction of increasing and decreasing air pollutants. To analyze the magnitude of trends, we used Sen's slope estimator, which is the robust estimator for the amplitude of trend slopes as proposed by Sen (1968):

$$\text{Slope} = \text{Median}\left(\frac{x_j - x_i}{j - i}\right) \quad (1 \leq i < j \leq n) \quad (7)$$

where slope is the monotonic increase or decrease rate, or the linear slope, of the entire data series  $x_k$  ( $k=1, 2, 3, \dots, n$ ) or any segmentation  $x_w$  ( $w=1, i+1, i+2, \dots, j$ ).

The PM<sub>2.5</sub> trend was calculated as the significant ( $P < 0.05$  in this study) slope of the Mann-Kendall (MK) trend test at each pixel's time series. Positive/negative trends were then defined as trends larger than zero/smaller than zero, respectively.

Median denotes the function to take the median value, and conducts a significance test on the result of the Sen's trend analysis using the MK approach (Wang et al., 2018).

### 3. RESULTS AND DISCUSSION

#### 3.1 Spatial Pattern of Average PM<sub>2.5</sub> Concentration

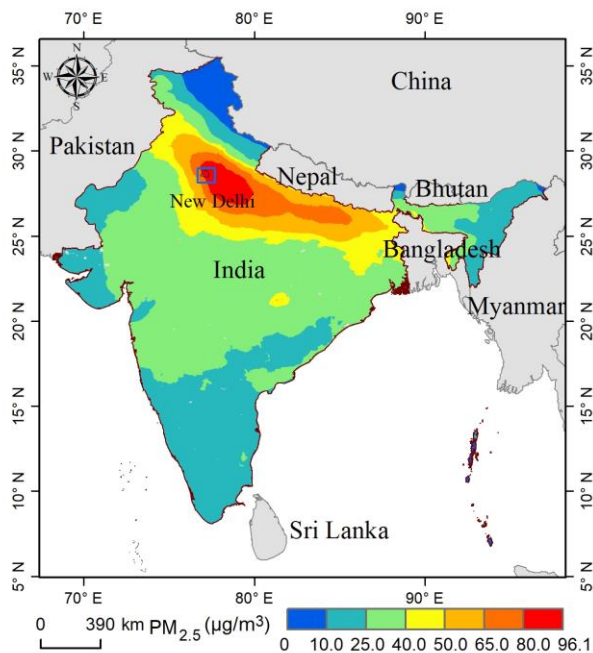


Figure 1. Spatial distribution of average PM<sub>2.5</sub> in India during 1998-2015

The annual mean of PM<sub>2.5</sub> from 1998 to 2015 in India presented an obvious three-stage stepped-down spatial distribution from the Ganges plain to north and south of India. Spatial distribution of each stage was distinctly different (Figure 1). High concentration of PM<sub>2.5</sub> was found in Ganges plain, which including the capital city—New Delhi. There are fertile land and excellent geographical conditions in the Ganges plain, which makes many cities with large population located, such as New Delhi, Kanpur and Lucknow. The air pollution caused by traffic exhaust is also one of the pollution sources that can not be ignored (Sharma et al., 2016), which brought great pressure to the local atmospheric environmental quality. In addition, Punjab and Haryana are India's principal agricultural areas, with a large proportion of farming and large-scale straw burning emissions. However, due to the Himalayas blocking, the region is prone to inversion and the overall atmospheric environment is relatively stable (Saikawa et al., 2019). To summarise, a large number of anthropogenic emissions of atmospheric pollutants, coupled with unfavorable conditions for the spread of atmospheric pollutants and the transmission of atmospheric pollutants from adjacent areas (e.g., Punjab, Haryana), which led to the Ganges plain becoming the most polluted area in India.

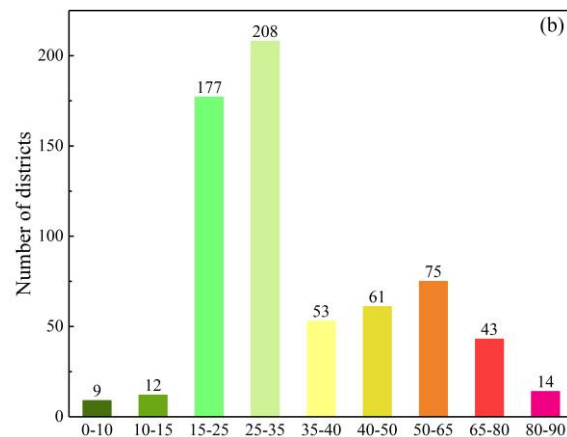
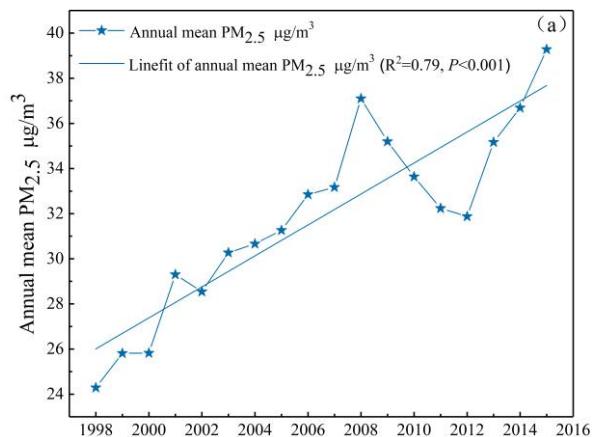


Figure 2. PM<sub>2.5</sub> annual mean change (a) and the number of districts different levels of annual mean PM<sub>2.5</sub> (b)

Figure 2a showed the annual mean PM<sub>2.5</sub> concentration change in India during 1998-2015. Overall, there was an upward trend of annual mean PM<sub>2.5</sub> concentration with  $R^2=0.79(P<0.001)$  in India. However, the change characteristics were variously in different stages. From 1998 to 2001, the average annual value of PM<sub>2.5</sub> showed a rapid upward trend (the average annual average of PM<sub>2.5</sub> in 1999-2000 was basically same), and reached a peak in 2001. In 2002, the annual mean PM<sub>2.5</sub> showed a brief and significant decline, but then it rebounded and continued to rise, and reached the highest value in 2008 since 1998. After that, the annual mean PM<sub>2.5</sub> experienced the longest period of continuous decline in the annual mean PM<sub>2.5</sub> in India during 1998-2015, while it dropped to valley value of  $32\mu\text{g}/\text{m}^3$

in 2012. After 2012, the annual mean PM<sub>2.5</sub> rose sharply and it reached nearly  $40\mu\text{g}/\text{m}^3$  in 2015, which reflected the air quality of India suffering a sharp deterioration tendency during 2012-2015. According to the number of districts in different annual mean PM<sub>2.5</sub> levels (Figure 2b), there are 9 districts with  $10\mu\text{g}/\text{m}^3$  meeting the air quality guidelines set by the World Health Organization. The number of districts reaching WHO IT-3 ( $15\mu\text{g}/\text{m}^3$ ), IT-2 ( $25\mu\text{g}/\text{m}^3$ ) and IT-1 ( $35\mu\text{g}/\text{m}^3$ ) is 12, 177 and 208, 1.82%, 26.86% and 31.56% of which accounts for the total districts respectively. Especially, 38.39% of the 659 districts with an annual average PM<sub>2.5</sub> were higher than  $40\mu\text{g}/\text{m}^3$ , indicating that more than one-third of district had experienced serious air pollution. In addition, there were 14

districts with an annual mean  $PM_{2.5}$  more than  $80\mu g/m^3$ , which showed that the air quality of those regions were extremely severe.

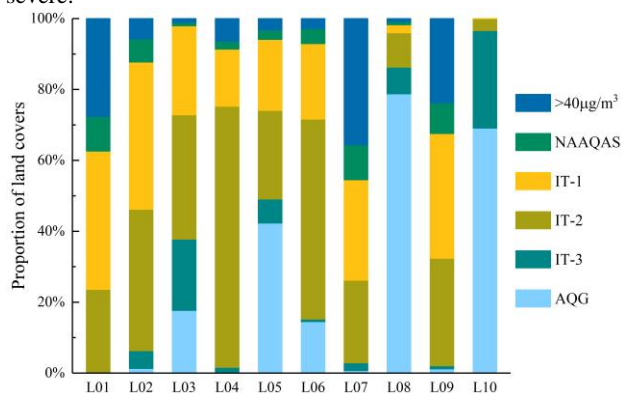


Figure 3. Proportion of  $PM_{2.5}$  at each level in different LC types

In order to further explore the characteristics of  $PM_{2.5}$  in different LC types in India, average  $PM_{2.5}$  from 1998 to 2015 in different LC types were counted respectively (Figure 3). Average  $PM_{2.5}$  were above  $40\mu g/m^3$  in the areas where the urban area (L07), farmland (L01) and water (L09) were the three largest types. The atmospheric pollutants caused by the mass population living in the urban area and the industrial pollutants caused by industrial production made the air pollution in L07 very severe. Different from  $PM_{2.5}$  in L07 and L09 types which occupied a larger proportion in NAAQAS level, annual average  $PM_{2.5}$  in shrub (L04), herbaceous cover (L03), woodland (L02) and sparse vegetation (L06) accounted for IT-1, IT-2 in most areas, indicating that air quality in these regions was relatively good. However, the air quality in these areas had not yet been optimistic and needed further attention. Grassland (L05), bare (L08) and permanent ice and snow (L10), where there was no obvious emission, were less affected by human activities. The  $PM_{2.5}$  in these regions was below  $10\mu g/m^3$  (AQG), and indicating that the air quality was the good.

### 3.2 Trend of $PM_{2.5}$ Concentration, 1998-2015

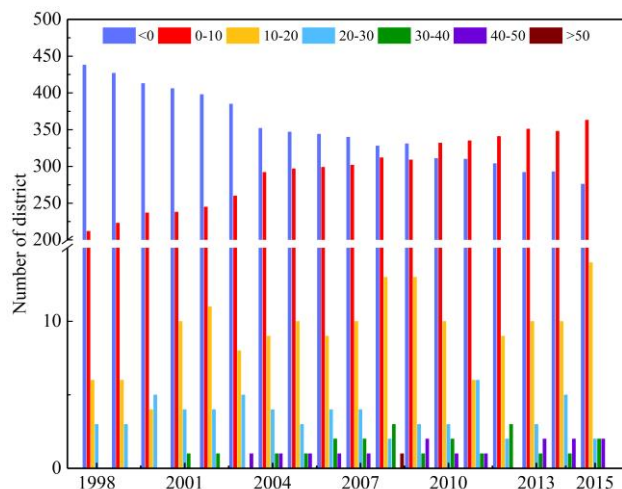


Figure 4. Districts of  $D_{PM_{2.5}}$  in India during 1998-2015

Figure 4 showed statistics of  $D_{PM_{2.5}}$  in 659 districts in India. The number of districts with  $D_{PM_{2.5}} > 0$  showed a significant increasing trend, especially the number of districts with  $D_{PM_{2.5}}$  between 0 and 10. However, the amount of districts with  $D_{PM_{2.5}}$

$< 0\mu g/m^3$  decreased from 438 in 1998 to 276 in 2015, indicating that the air pollution in Indian districts was getting worse. The amount of districts with  $D_{PM_{2.5}}$  between 10 to  $20\mu g/m^3$  was more than doubled from 6 in 1998 to 14 in 2015. In addition, districts with  $D_{PM_{2.5}} > 30\mu g/m^3$  emerged continuously, especially in 2003, it overpassed  $40\mu g/m^3$ , indicating the air quality situation in urban areas had further deteriorated.

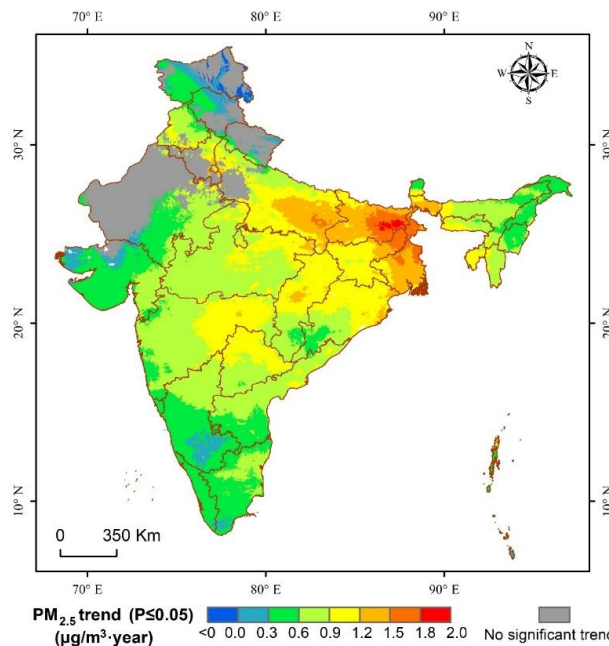


Figure 5. Significant trends of  $PM_{2.5}$  using MK test and Sen's slope in India during 1998 to 2015

The area with strong significant positive trends was obtained mainly in the Ganges plain, northern Chhattisgarh and Jharkhand. In particular,  $PM_{2.5}$  in Patnabang and its adjacent West Bengal region located in the southeastern part of the Ganges River region was increasing at a rate of  $1.2\mu g/m^3$  or more. While the annual growth trend of  $PM_{2.5}$  in most of southern India (including Tamil Nadu, Kerala, Andhra Pradesh, Goa and Karnataka), and western India (including Gujarat and western Rajasthan) were less than  $0.9\mu g/m^3$ . There was no significant change of  $PM_{2.5}$  in the western Thar Desert and the Himalayas area in the northwest of India, and the  $PM_{2.5}$  even had a significant downward trend in the partial areas of Himalayas. The main LC types in these areas were permanent ice and snow and water, which was likely due to the low intensity of human activities and the low emissions of air pollutants.

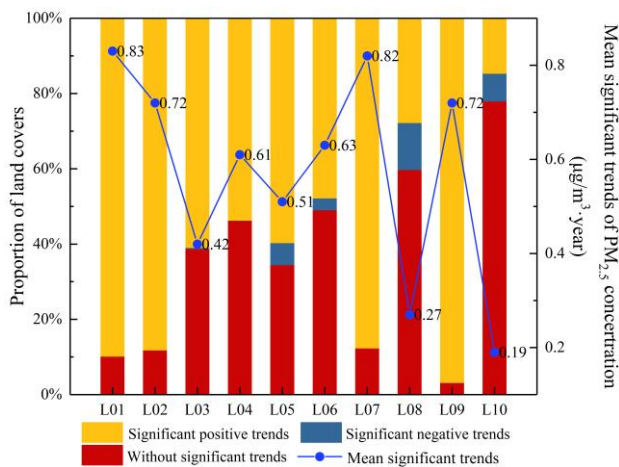


Figure 6. Significant trends of PM<sub>2.5</sub> in different LC types during 1998-2015

We can see from the Figure 6 that except for bare area and permanent ice and snow LC types, the other types of PM<sub>2.5</sub> had a more proportion significant positive trend than others, indicating that PM<sub>2.5</sub> of these types was mainly increasing. The average increase trend of farmland type (0.83µg/m<sup>3</sup>) was the largest in all LC types, followed by the urban area (0.82µg/m<sup>3</sup>). The average annual growth rate of PM<sub>2.5</sub> in forest and water LC types is stable, but the proportion of significant positive trends in water is obviously higher than that in forest, indicating that the water is suffering from more serious air pollution than forest.

#### 4. CONCLUSIONS

In this study, we utilized PM<sub>2.5</sub> and LC products from remote sensing first analysed the differences in PM<sub>2.5</sub> concentrations between urban and surrounding areas in India, and further investigated the response of PM<sub>2.5</sub> to LC using MK test and Sen's slope methods during 1998 to 2015. The following conclusions can be drawn:

- 1) The number of districts with  $D_{PM_{2.5}} > 0$ , especially those with between 0 and 10, increased continuously during 1998-2015. The number of districts with  $D_{PM_{2.5}} > 0$  predominated since 2010.
- 2) The most serious PM<sub>2.5</sub> pollution areas were located in the northern India, especially Ganges plain. The annual average PM<sub>2.5</sub> showed an upward trend from 1998 to 2015. Areas of annual mean PM<sub>2.5</sub> higher than NAAQAS were mainly distributed in farmland and urban areas, which accounted for 58% of India's area. However, there was only 10% of areas with  $PM_{2.5} < 10\mu g/m^3$ , and these areas were mainly found in the permanent ice and snow (L10) in northwestern India.
- 3) 90% of the cropland and 88% of the urban area were found significant positive trends and the average trends of PM<sub>2.5</sub> in these regions were 0.83µg/m<sup>3</sup> and 0.82µg/m<sup>3</sup> respectively, which were higher than those other land covers.

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